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The potential health challenges of TiO₂ nanomaterials

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ABSTRACT: Titanium dioxide (TiO₂) nanomaterials (NMs) have found widespread applications owing to their attractive physical and chemical properties. As a result, the potential adverse impacts of nano-TiO₂ exposure on humans have become a matter of concern. This review presents the state-of-the-art advances on the investigations of the adverse effects of NMs, including the potential exposure routes of nano-TiO₂ (e.g. respiratory system, skin absorption and digestive system), the physico-chemical characterizations of nano-TiO₂ (e.g. crystal structure, shape,size, zeta potential, treatment media, aggregation and agglomeration tendency, surface characteristics and coatings), risk evaluation of nanotoxicity (e.g. cytotoxicity, ecotoxicity, phototoxicity, and phytotoxicity) and potential mechanisms of adverse effects (e.g. generation of reactive oxygen species, oxidative stress and organelle dysfunction). The review aims to facilitate scientific assessments of health risks to nano-TiO₂, which would guide the safe applications of NMs in our daily life. Copyright © 2015 John Wiley & Sons, Ltd.

Keywords: titanium dioxide; nanomaterials; health risk; route; characterization; mechanism

Introduction

With increasing exposure of dispersed nanomaterials (NMs) from products or workplaces, there is an increased opportunity for NMs to enter the body, leading to potential adverse effects. For example, two female patients, who worked in a printing plant, died of respiratory failure (Song et al., 2009). Pathological examinations have aroused concern that long-term exposure to NMs without protective measures may lead to serious damage to human lungs (Song et al., 2009). Owing to the lack of systematic knowledge of toxicology, epidemiology and workplace exposure of NMs (Kuempel et al., 2007), many researchers doubt the adverse effects of NMs (Gilbert, 2009; Brain et al., 2010), although the disaster triggered by asbestos dust exposure during the past decades is still fresh in the mind (Frost et al., 2008). Recently, Hristozov et al. (2014) suggested that the hazard identification of NMs under multiple regulatory frameworks is needed, which should integrate individual studies about physico-chemical and toxicological properties of NMs. Schulte et al. (2008) reviewed a conceptual framework using occupational risk protection of engineered NMs, including the potential control of the different routes and adverse factors of NMs exposure. To be doubly sure, there is an urgent need for a better understanding of the risk of emerging NMs [e.g. nanoscale titanium dioxide (nano-TiO₂)] before they are put into practice.

Nano-titanium dioxide (TiO₂), as the most prevalent engineered NMs, is widely used in cosmetics, orthodontic composite, chewing gum, wastewater disinfectant, antibacterial, additive in pharmaceuticals, etc (Chen *et al.*, 2013; Heravi *et al.*, 2013; Prasad *et al.*, 2013; Yang *et al.*, 2013). Specifically, sunscreen is a \$400 million dollar industry, 70% of which contains nano-TiO₂ (Nel *et al.*, 2006). Furthermore, nano-TiO₂ led to ecotoxicity, phytotoxicity, cytotoxicity and genotoxicity in many researches (Diana *et al.*, 2010; Gottschalk *et al.*, 2013; Zhao *et al.*, 2013; Josko and Oleszczuk 2014). Unfortunately, most of the existing literature about nano-TiO₂ focused on the preparation methods of NMs, and many risk assessment studies showed shortcomings in experimental design (Warheit, 2013; Hristozov *et al.*, 2014; Wang *et al.*, 2014). Up until

now, reliable methods to identify relevant health risks of various forms of nano-TiO₂ have not been proposed.

Based on the abundance of risk data generated on nano-TiO₂, four parts are included in this review to facilitate the scientific risk assessments of nano-TiO₂ exposures. (1) Exposure routes of NMs: nano-TiO₂ can interact with the body through various routes, which should be well known for the reasonable design of experiment and risk evaluation. (2) The characterization of nano-TiO₂: different types of nano-TiO₂ have a variable toxicity potential, making it necessary to obtain adequate material characterization for the interpretation of measured results. (3) Risk evaluation in different species: various species have been employed to simulate human exposures, which can help to gauge the relevance of health risks for human beings. (4) The potential mechanisms of nanotoxicity: several major mechanisms have been proposed for inducing the adverse effect of nano-TiO₂, which would guide the usage of NMs in daily life and obtain the benefits of nanotechnology under safe situations.

Exposure route of nano-TiO₂

Various exposure routes of nano- TiO_2 entering the human body through the respiratory system, skin and digestive system have

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^dThe Key Laboratory of Biomedical Information Engineering of Ministry of Education, School of Life Science and Technology, Xi'an Jiaotong University, Xi'an 710049, China Health risk of nano-TiO₂

Table 1. Risk evaluat	ion of nano-TiO ₂ in	human cells				
Tissues/Cell	Crystal structure	Shape	Size (nm)	Treatment media	Zeta potential (mV)	Aggregation
Human embryonic kidney	Anatase	Spherical	22.9±0.3 (21), 50.7±0.4 (50)	Water	8.71 mV (21 nm), 9.38 mV (50 nm)	No
Human Caco-2 _{BBe1} cells	Anatase	Spherical	336±6, 365±5	Cell culture medium	Negative charge above pH=4	Yes
Human erythrocyte and lymphocyte	Anatase/rutile	Spherical	35–56	PBS	-	Yes
Lung fibroblast cells	Rutile/ anatase	Spherical	40–200	-	-	Yes
Dermal fibroblasts	Anatase	-	15	Cell culture medium	-	Yes
Bronchial epithelial	Rutile Apatase	Needle-like Spherical	10 × 40	Growth media	-	Yes
Lung fibroblasts (IMR-90)	Anatase	Spherical	<100	Water	+48.8	Yes
Bronchial epithelial cells	Anatase	-	53, 311 461 86, 356	Water PBS Cell culture media	-24 -11 -9	Yes
Macrophage-like human THP-1 cells	Anatase	Spherical	<50 <25 10	-	-	-
	Rutile	Creigulau	<5 30–40			
Human keratinocytes	-	Spherical	414.9±4.5	Cell culture media	-27.9±0.3	Yes
Keratinocyte HaCaT cells	Anatase	Granular	173.8, 259, 263.9	Cell culture media	-	Yes
Lymphocytes	-	Spherical	90-110	-	-	No

been found. As the production of nanoindustry, nano-TiO₂ may be easily inhaled by workers after respiratory system exposure, leading to the lung damage. The concentration of total airborne TiO₂ particle in the bin filling area of the facility ranges from 15 000 to 156 000 particles/cm³, and more than 97% of them were less than 100 nm (Berges et al., 2007). An epidemiological study demonstrated that NMs-exposed workers from six European countries showed a higher incidence of lung cancer than general populations (Liao et al., 2008; Wang et al., 2009). As a result, the permissible exposure limit and the immediately dangerous concentration of nano-TiO₂ for worker safety were estimated as 15 mg/m³ and 7 500 mg/m³, respectively (Liao et al., 2008; NIOSH 2011; Wang et al., 2009). For the populations heavily exposed to nano-TiO₂, the respiratory system is the major exposure route.

Currently, there is increased usage of nano-TiO₂ in cosmetics industry because of its absorptive properties (Singh et al., 2009), although people argue that the marketing of nano-TiO₂ is ethically undesirable (Jacobs et al., 2010). With the cumulative exposure during our daily life, the interaction between dispersed nano-TiO₂ from cosmetics products and the skin are unavoidable. The amounts of nano-TiO₂ applied on the skin would be 8-37 mg kg⁻¹ weight for an adult, assuming that a sunscreen contains 5% nano-TiO₂ (Davis et al., 2010). Wu et al. (2009) found that nano-TiO₂ can penetrate through the hairless mice skin after 60-days exposure, inducing diverse pathological lesions, especially in skin and liver. In addition, nano-TiO₂ in a sunscreen penetrated into deeper areas of the stratum corneum in psoriatic skin than healthy skin, because of the looser

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Table 1. (Continue	d)			
Tissues/Cell	Concentration	Exposure duration	Toxicity	Reference
Human embryonic kidpov	10–1 000 μg/mL	1 hr, 3 weeks	DNA damage, cell-transformation and cell- anchorage independent growth in soft-	(Abrusci <i>et al.,</i> 2013)
Human Caco-2 _{BBe1} cells	0.35–35 μg/mL	24 hrs	Food grade nano-TiO ₂ exposure led to the loss of microvilli from the Caco-2BBe1 cell system.	(Archana <i>et al</i> ., 2013)
Human erythrocyte and lymphocyte	25–500 μg/mL	3 hrs	Led to the decreased activity of mitochondrial dehydrogenase, DNA damage, and apoptosis in lymphocyte; interacted with hemoglobin and showed hemolysis in erythrocyte.	(Ghosh <i>et al.</i> , 2013)
Lung fibroblast cells	25–400 μg/mL	24, 48 hrs	Intracellular ROS, loss of mitochondrial membrane potential, cell cycle progression alterations, TNF and CYP1A gene expression change were shown after exposure.	(Drevet <i>et al.</i> , 2012)
Dermal fibroblasts	1–100 μg/mL	24 hrs	Increased the phosphorylation of H2AX, ATM, and Chk2, and inhibited DNA synthesis and frequency of replicon initiation.	(Vyas <i>et al.</i> , 2011)
Bronchial epithelial cells (BEAS-2B)	1–100 μg/cm ²	24, 48, 72 hrs	DNA damage Increased micronuclei	(Falck <i>et al.</i> , 2009)
Lung fibroblasts (IMR-90)	$2-50 \ \mu\text{g/cm}^2$	2, 4, 6, 24 hrs	DNA adduct formation	(Bhattacharya <i>et al</i> ., 2009)
Bronchial epithelial cells	10–40 μg/cm ²	0.5, 1, 2, 4, 24 hrs	Induced lipid peroxidation, lysosomal membrane destabilization, and cell death.	(Hussain <i>et al.,</i> 2010)
Macrophage- like human THP-1 cells	20–500 μg/mL	24 hrs	Smaller anatase and larger rutile NMs provoked higher IL-1β production. Spicular NMs induced higher toxicity than similar sized and spherical NMs.	(Morishige <i>et al.</i> , 2010)
Human keratinocytes	0.5–10 μg/mL	3 months	Led to changes of cell cycle and the increase of apoptotic cells.	(Kocbek <i>et al.</i> , 2010)
Keratinocyte HaCaT cells	47.0–60.2 μg/mL	24 hrs	Affect cell-matrix adhesion	(Fujita <i>et al.</i> , 2009)
Lymphocytes	0.25–2 mM	3, 6, 24 hrs	Genotoxicity and cytotoxicity	(Ghosh <i>et al.</i> , 2010)

corneocyte organization in the diseased human skin (Pinheiro et al., 2007).

Nano-TiO₂ have been identified in food-grade TiO₂ (Yang *et al.*, 2014) and preferred for drinking water treatment, making them highly possible to be absorbed by digestive system in the process of feeding, and redistributing to other tissues/organs. For instance, Nano-TiO₂ accumulated in mice liver, spleen, kidney, lung and brain after a single oral gavage of 5 000 mg kg⁻¹ (Wang *et al.*, 2007). Furthermore, nano-TiO₂ can spread to the environment owing to industrial processes and consumer products, such as the use of wastewater disinfectant (Yang *et al.*, 2013). Nano-TiO₂ present in the environment can affect the water, soil and related organisms, leading to high mortality of medaka (Ma *et al.*, 2012), alteration of bacterial community composition and function in ecosystems

(Ge *et al.*, 2011; Binh *et al.*, 2014), and changes in germination and root elongation in plants (Song *et al.*, 2013a, 2013b). These contaminations will finally induce an adverse effect on humans through the digestive system after feeding (e.g. water and vegetables), because humans are at the top of the food chain (D'Agata *et al.*, 2014).

The respiratory system, skin and digestive system are three major routes that nano-TiO₂ can enter the human body to induce potential health risk. Other routes of nanotoxicity assessment, such as intravenous injection, intraperitoneal injection and implantation, have also developed and widely used in animal testing (Patra *et al.*, 2009; Yang *et al.*, 2009). Although the relevance of animal experiments used to simulate human exposures is still under debate, these methods are regarded as part of the above three routes.

Characterization of nano-TiO₂

NMs can penetrate cells with no need of specific receptors on the outer surface of the cells, which may be attributed to a passive uptake or adhesive interaction through van der Waals forces, electrostatic charges or steric interactions (Rimai *et al.*, 2000; Peters *et al.*, 2006). To provide an objective evaluation of health risk after nano-TiO₂ exposures, one of the important things is to obtain the adequate material characterization of test samples (Warheit, 2013). Recent studies suggested that the physico-chemical properties of

nano-TiO₂ affect the result interpretation *in vitro* and *in vivo* after NMs exposure, such as crystal structure, shape, size, zeta potential, treatment media of NMs, aggregation and agglomeration tendency, surface characteristics and coatings (Scherbart *et al.*, 2011; Tyner *et al.*, 2011; Silva *et al.*, 2013).

Crystal structure

X-ray diffraction (XRD) is the basic tool to analyze the crystal structure of nano-TiO₂ (Chen *et al.*, 2013). Anatase, rutile and

Table 2. Risk ev	aluation of nano-Ti	iO ₂ in mice tis	sues and cel	S			
Tissues/Cell	Crystal structure	Shape	Size (nm)	Treatment media	Zeta potential (mV)	Aggregation	Surface modification
Reproductive system	Anatase	-	208–330	HPMC + K4M	9.28 mV	Yes	-
Airway	Anatase/rutile	-	21	PBS	-	Yes	-
Sperm	Anatase	-	25	PBS with 0.5 % Tween80	-	-	-
Kidney	-	-	21	Water	-	Yes	-
Luna	Rutile	Elongated	12.8–486	Air	-	Yes	Zr, Si, Al
Lung	Anatase/rutile	-	200–300	PBS + DPPC + BSA	-	Yes	-
Lung	Anatase	-	80–110	-	-	Yes	-
Brain	Anatase	-	5	HPMC + K4M	-	-	-
Brain	Anatase	-	2 570	Saline	-	-	-
Brain	Anatase	-	5	HPMC + K4M	-	-	-
JB6 cell line	Anatase/rutile	-	34.9±16.8	PBS	-	-	-
L929 fibroblast cell	Anatase	-	20–50	Culture medium	-	Yes	-
Skin Kidney Liver	Anatase Anatase Anatase/rutile Rutile Rutile Rutile	-	4 10 21 25 60 90	Caprylic/ LCC + Tween 80 + carbopol + triethanolamine + water	-	-	-
Liver	Anatase	-	5	HPMC + K4M	-	-	-
Offspring	Rutile	Elongated	12.8–486	Air	-	Yes	Zr, Si, Al

PBS, phosphate buffered saline; DPPC, dipalmitoyl phosphatidylcholine; BSA, bovine serum albumin; i.p., intraperitoneal administration; i.g., intragastric administration; HPMC, hydroxypropylmethylcellulose; SC, subcutaneous; LCC, capric triglyceride, SOD, superoxide dismutase; MDA, malondialdehyde. brookite are the three major crystal structures for nano-TiO₂, and the first two are the most common forms in toxicity assessment of nano-TiO₂ because brookite is rare (Warheit, 2013). One study suggested that anatase nano-TiO₂ showed more toxicity in human lung epithelial cells (A549) than rutile nano-TiO₂ (Sayes *et al.*, 2006). In contrast, the anatase/rutile nano-TiO₂ mixture showed a higher mortality of Zebrafish larvae than anatase nano-TiO₂ under ultraviolet (UV) irradiation, because the crystal structure affected the photocatalytic properties of nano-TiO₂ (Clemente *et al.*, 2014).

Shape, size and specific surface area

With the changes in synthesis condition (e.g. raw material, temperature, acidic and alkaline conditions), nano-TiO₂ with various shapes (e.g. rods, dots and belts), sizes and specific surface areas have been prepared for different applications (Wang *et al.*, 2004; Barnard and Curtiss, 2005). Nano-TiO₂ have structural components smaller than 100 nm in at least one dimension (Cristina *et al.*, 2007), which need transmission electron microscope (TEM) and scanning

Table 2. (Conti	nued)			
Tissues/Cell	Concentration	Exposure duration	Toxicity	Reference
Reproductive system	2.5–10 mg/kg	i.g., 90 days	Induced fertility reduction and ovary injury through alteration of inflammation/ follicular atresia-related cytokine expressions.	(Zhao <i>et al.,</i> 2013)
Airway	193±8 mg/lung after eight repeated exposures	Inhalation	Increased infiltration of neutrophils in airways of naïve mice, leading to neutrophilic airway inflammation and the loss of weight in asthmatic mice.	(Jing <i>et al.</i> , 2014)
Sperm	10–250 mg/kg	Oral, 42 days	Reduced the serum T level and spermatogenesis, inducing the spermatozoa abnormality in epididymides.	(Jia <i>et al.,</i> 2014)
Kidney	0.1, 0.25, and 0.5 mg/week	i.g., 4 weeks	Led to renal fibrosis through a ROS/RNS- related HIF-1alpha-upregulated TGF-beta signaling pathway.	(Kim <i>et al.,</i> 2013)
Lung Lung	42 mg/m³ 40 mg/kg	Inhalation, 1 hr/day i.p.	Induced long-term lung inflammation. Induced acute lung inflammation through oxidant-dependent inflammatory signaling and NF-kB pathway.	(Hougaard <i>et al.</i> , 2010) (Moon <i>et al.</i> , 2010)
Lung	324–2 592 mg/kg	i.p.,7, 14 days	Induced alveolar septal thickening, neutrophil infiltration, thrombosis.	(Chen <i>et al</i> ., 2009)
Brain	5-50 mg/kg	i.g., 60 days	Impaired behaviors of spatial recognition memory, disturbing homeostasis of trace elements, enzymes, and neurotransmitter systems.	(Hu <i>et al.</i> , 2010)
Brain	1 μg/μL	SC injection on gestational days 6, 9, 12, 15	Changed the gene expression associated with brain development, cell death, oxidative stress, mitochondria, inflammation, and neurotransmitters.	(Shimizu <i>et al.,</i> 2009)
Brain	5–150 mg/kg	i.p., 14 days	Translocated into brain, leading to OS, shape changes and inflammation in neurons.	(Ma et <i>al.</i> , 2010)
JB6 cell line	1–20 μg/cm ²	24, 48, 72 hrs	Induced the activation of mitochondrial- related pathways and cell apoptosis.	(Zhao <i>et al.</i> , 2009)
L929 fibroblast cell	0–300 mg/mL	48 hrs	Suppressed carbohydrate metabolism in L929 cells.	(Jin <i>et al.</i> , 2013)
Skin Kidney Liver	400 μg/cm ²	Skin exposure, 60 days	Changed the product of MDA and reduced the content of HYP in skin; saccus lymphaticus was shown in kidney; induced focal necrosis and liquefaction necrosis in the liver.	(Wu <i>et al.,</i> 2009)
Liver	5–50 mg/kg	i.g., 60 days	Accumulated in liver, inducing histopathological changes, hepatocytes apoptosis, and dysfunction of mice liver.	(Cui <i>et al.</i> , 2011)
Offspring	42 mg/m ³	Inhalation, 1 hr/day	Moderate neurobehavioral alterations	(Hougaard <i>et al.</i> , 2010)

Table 3. Risk evalue	ation of nano-TiO	12 in rat tissu	ies and cells							
Tissues/Cell	Crystal structure	Shape	Size (nm)	Treatment media	Zeta A potential (mV)	ggregation	Concentration	Exposure duration	Toxicity	Reference
Alveolar macrophages	Anatase	1	< 100		1	1	2.5–100, µg/mL	24 hrs	Cellular response of macrophages to nano-TiO ₂ is age dependent. The generation of O_2^{-1} and NO increased in a dose-dependent manner, IL-10 release was shown in the highest exposure of nano-TiO ₂ .	(Gu <i>et al.,</i> 2013)
Liver, lung, kidney and blood cells	Anatase	Spherical, lentil-like	5, 10,150	0.9% Nacl	ı		16 g/kg	ip, 3, 12 months	Nano-TiO_ accumulated inside mononuclear cells and in liver and lung at 3 and 12 months, with higher reactive and biopersistent in lung, and liver cell membrane damage.	(Kunze <i>et al.</i> , 2009)
Neuroglia/ brain		ı	10,20 and 200	MEM	ı	ı	0.1,1 and 10 mg/kg	Inhalation, 72 hrs	Nano-TiO ₂ induced the change of IL-1β, TNF-a and IL-10 levels, blood-brain barrier damage, necrosis, inflammation and cellular oedema in brain.	(Alammar <i>et al.</i> , 2013)
PC12	Anatase	I	24-697	Cell culture media	ı	Yes	1–100 µg/mL	6, 12, 24, 48 hrs	Induced a time- and dose- dependent cytotoxicity, intracellular accumulation of ROS, and cell apoptosis.	(Chusuei <i>et al.</i> , 2013)
Lung	Anatase/rutile	ı	100	Aerosol		Yes	6 mg/m³	Inhalation, 4, 12, 24 hrs	Led to increased microvascular OS and nitrosative stress, decreased NO bioavailability, and microvascular distinction	(Nurkiewicz <i>et al.</i> , 2009)
Lung	Anatase	Belt	1 536 ± 139	Dispersion . media	-30.3 ± 2.8	Yes	20–200 µg/rat	i.t. instilled, 1 day	Induced significant inflammation in BALF, NMs had not cleared from alveolar macrophages by 7 days and recovered from the lung.	(Silva <i>et al.</i> , 2013)
Lung	·	ı	5-200	ı	ı	·	0.5–50 mg/kg	i.t., 1 day	Damage cell structure and led to pulmonary alveolar macrophage dvsfunction.	(Liu <i>et al.,</i> 2010)
i.p., intraperitoneal distearoylphosphatic	administration; i: Jylcholine + low-	t., intratrach endotoxin ra	neal instillativ at serum albı	on; BALF, bı umin)	ronchoalveo	lar lavage	fluid; ROS, reac	ctive oxygen	species; dispersion medium (etha	anol + sterile saline +

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		Reference	(Davis <i>et al.</i> , 2010)				(Wang <i>et al.</i> , 2005)						(Ma <i>et al.</i> , 2012)						(Griffitt et al., 2009)
		Toxicity	TiO ₂ inhibited the growth	C brevis and S. costatum	lipid oxidation inside	al cells.	TiO ₂ formed clusters	ר the algae cells, affecting	transfer of energy	ughout depleting the	centration of available	d for Daphnia magna.	ed dose-dependent	rtality in <i>medaka</i> , and	rrs LC50 under laboratory	t and solar UV radiation	e 294 and 2.46 mg/L,	oectively.	jed gene expressions on

Table 4. Risk evaluati	on of nand	o-TiO ₂ in a	quatic animal	S							
Species	Crystal structure	Shape	Size (nm)	Treatment media	Zeta / potential (mV)	Aggregation	Surface modification	Concentration	Exposure duration	Toxicity	Reference
Karenia brevis and Skeletonema costatum	Anatase	Granular	<40	F/2 medium	1	Yes	1	125–750 mg/L	24, 48, 72 hrs	Nano-TIO ₂ inhibited the growth of <i>K. brevis</i> and <i>S. costatum</i> through the ROS production and lipid oxidation inside	(Davis <i>et al.</i> , 2010)
Daphnia magna	Rutile	Spherical	20	Water		Yes	,	1 and 10 mg/L	1–24 days	Nation 2012, formed clusters with the algae cells, affecting the transfer of energy throughout depleting the concentration of available food for <i>Dathnia maana</i> .	(Wang <i>et al</i> , 2005)
Japanese medaka	Anatase/ rutile	,	22-121	Water	16.5	Yes		2-7 mg/L	4 hrs	Showed dose-dependent mortality in <i>medaka</i> , and 96 hrs LC50 under laboratory light and solar UV radiation were 294 and 2.46 mg/L, respectivelv.	(Ma <i>et al.</i> , 2012)
Zebrafish (Gills)		Spherical	220.8, 687.5	Water	-25.1	Yes		1 000 µg/L	2–48 hrs	Changed gene expressions on ribosome structure and activity.	(Griffitt <i>et al.</i> , 2009)
Zebrafish (Embryo)	Anatase	Spherical	>50	Water	I	Yes	G	10, 20 ppt	2–72 hrs	Nano-TIO ₂ penetrated into embryo and led to damaged eyes, abnormal heart and chord formation, increased activities of GSH, CAT, GST, and GSR.	(Yeo and Kang 2009)
Daphnia magna	Anatase/ rutile		580.5, 2 349, 3 528.6	Water	,	Yes		0.1–100 mg/L 0.1–5 mg/L	24, 48, 72 hrs 21 days	Showed minimal and high toxicity after 48 and 72 hrs exposures to <i>daphnia</i> . 21days accumulation induced severe growth retardation, mortality and	(Zhu et al., 2010)
GSH, glutathione; CAT,	catalase; (GST, glutat	thione S-trans	ferase; GSR, ç	glutathione	e reductase.				ובהוחמתרוואב מבוברוס.	

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Table 5. Risk ev	aluation of nano-	-TiO ₂ in b	acteria								
Species	Crystal structure	Shape	Size (nm)	Treatment media	Zeta potential (mV)	Aggregation	Surface modification	Concentration	Exposure duration	Toxicity	Reference
B. subtilis/ A. hydrophila	Anatase Anatase/rutile	·	300–500 22.9	Water	1	Yes -	1	0.5-25 mg/L	1-2 hrs	Led to the reduction of viable <i>B. subtilis</i> and <i>A. hydrophila</i> under simulated	(Binh <i>et al.</i> , 2014)
Soil Bacterial Communities	Anatase/rutile	ı	194	Water	I	Yes	I	20 mg/g	288 days	Changed the bacterial community composition and reduced diversity throuch soil water	(Ge <i>et al.</i> , 2013)
B. subtilis/E. coli	Anatase/rutile	1	320-330	Water	1	Yes	ı	10-5 000 ppm	20-26 hrs	Induced a dose- dependent growth reduction in bacteria, such as 2000 ppm NMs led to 99% and 46% growth reduction in <i>E. coli</i> and <i>B. subtilis</i> ,	(Adams <i>et al.</i> , 2006)
B. cereus 905	Anatase/rutile		21	Water	T			0.002-2 mg/mL	12 hrs	Changed community structure and reduced bacterial population.	(Wang <i>et al.</i> , 2009)

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Table 6. Risk eval	luation of nano-Ti	iO ₂ in plants								
Species	Crystal structure	Shape	Size (nm)	Treatment media	Zeta potential (mV)	Aggregation	Concentration	Exposure duration	Toxicity	Reference
Lepidium sativum	I	- I	<21	Water	-7.30		10–1 000 mg/L	72 hrs	Showed more toxic in water than in soil and inhibited root	(Josko and Oleszczuk 2014)
Anabaena sp	Rutile	Spherical	450-650	Culture media	- 7.8	Yes	1–50 mg/L	96 hrs	Induced toxicity at concentrations of 10-50 mg/L. Low dose (<1 mg/L) of NMs significantly enhanced the toxicity of Zn^{2+} . Toxicity of Zn^{2+} . nano-TiO ₂ system reduced with	(Tang <i>et al.</i> , 2013)
Chlamydomo- nas reinhardtii	Anatase	Spherical	Ŋ	Exposure media	-1.89±0.19μm cmV ⁻¹ s ⁻¹	Yes	1 mg/L	2 hrs	increasing NMs doses. Decreased functions of photosynthesis-related genes, and increased levels of proteasome transcripts encoding	(Simon et al., 2013)
Marine phytoplankton	Anatase/rutile	Semi-spherical	15–30	Seawater	ſ	Yes	1–7 mg/L	96 hrs	under increased ROS through UV radiation and show toxicity to	(Miller <i>et al.</i> , 2012)
Lactuca sativa	Anatase/rutile		27	Water	ſ	Yes	100–5 000 mg/L	2-15 days	Took up into plant seeds, affecting root elongation in <i>L. sativa</i> .	(Song <i>et al.</i> , 2013)

electron microscope (SEM) to image their morphology (Warheit et al., 2006; Chen et al., 2013). As shown in the published literature, the changes in morphology resulted in the different adverse effects after nano-TiO₂ exposure. For example, compared with TiO₂ nanospheres (diameter 60–200 nm) and short TiO₂ nanobelts $(0.8-4 \mu m \log, 60-300 nm wide)$, $\log TiO_2$ nanobelts $(15-30 \mu m m)$ long, 60-300 nm wide) induced significantly higher cytotoxicity in alveolar macrophage after 4-h exposure at dosages of 100 and 200 μ g ml⁻¹ (Hamilton *et al.*, 2009). Another study showed that spicular nano-TiO₂ stimulated a higher production of Interleukin (IL)-1 β than spherical nano-TiO₂ at lower concentrations (20 and 100 μ g ml⁻¹), although they had a similar size and identical rutile structure (Morishige et al., 2010). Park et al. (2014) suggested that anatase nano-TiO₂ [180.2 \pm 50.8 nm and –11.15 \pm 0.46 mV in fetal bovine serum (FBS)] showed a higher accumulation in the heart, lung and liver of mice than brookite nano-TiO₂ (126.8 \pm 36.5 nm and -9.66 ± 0.42 mV in FBS), although they had the same rod shape. Besides shape, size is another main determining factor for the distribution of nano-TiO₂ after respiratory tract exposure (Oberdorster et al., 2004). For example, amongst the nano-TiO₂ of three different sizes (5, 21 and 50 nm) with a dose more than 5 mg kg⁻¹, 5 nm nano-TiO₂ induced the most severe pulmonary toxicity in rats (Liu et al., 2009). The difference in the shape and size also changes the specific surface areas of nano-TiO₂, where a larger specific surface area had been found to induce higher adverse effects in vitro and in vivo after incubation, especially for the specific surface area-dependent phototoxic effect (Xiong et al., 2013; Park et al., 2014).

Treatment media

During the whole life-cycle of a product, the pristine characterizations of nano-TiO₂ may be changed during the use and disposal phases (Al-Kattan et al., 2014). Compared with the current focus on unaltered NMs, it is necessary to determine the behavior of nano-TiO₂ under a practical situation (e.g. treatment with media) before risk assessment. For instance, the same nano-TiO₂ gave the sizes of 53-311 nm, 461 nm and 86-356 nm in water, PBS and cell culture media, respectively (Hussain et al., 2010). The cultured cells have been used widely as the model system for cytotoxicity assessment in vitro. However, Prasad et al. (2013) found that the treatment media of NMs affected particle uptake, cell-NMs interaction and chromosomal damage through facilitating particle agglomeration. Additives (e.g. bisphenol A) in cell culture media led to synergistic toxicity with NMs through increasing the aggregation level and zeta potential of nano-TiO₂ (Zheng et al., 2012). Stomach acid (pH<4) can change the zeta potential of food-grade nano-TiO₂ after intragastric (i.g.) administration, which differ from the test results around pH 7 in deionized water (Yang et al., 2014). As a result, the situations concerning nano-TiO₂ should be paid more attention in nanotoxicity assessment (Karunakaran et al., 2013).

Modification

Nano-TiO₂ has the most exclusive absorption of UV, which corresponds to 3-5% of solar irradiation. Many modifications are focused on obtaining nano-TiO₂ activity in the range of visible light for the use of solar radiation, such as Fe-doped TiO₂ nanorods, N-doped TiO₂ nanoparticles, phosphate-modified nano-TiO₂ and polymers modified nano-TiO₂ (Liu *et al.*, 2011; Chakrabortty and Gupta 2013). The modified nano-TiO₂ can degrade soluble organic

pollutants in water through photocatalysis (Orlov *et al.*, 2006), in the meantime, nano-TiO₂ with modification (trap centre) can eliminate the generation of free radical under solar energy, differing from the inducing effect of nano-TiO₂ in free radical generation (Wakefield *et al.*, 2004).

Risk evaluation of nanotoxicity

Based on the inevitable existence in air, water, soils and organisms, the potential toxicity of nano-TiO₂ to health and the environment has been investigated at cell, tissue and animal levels, including human cells (Table 1), mice (Table 2), rats (Table 3), aquatic animals (Table 4), bacterium (Table 5) and plants (Table 6). Cytotoxicity, ecotoxicity, phototoxicity and phytotoxicity have been found after nano-TiO₂ exposure (Bhattacharya *et al.*, 2009; Miller *et al.*, 2012; Binh *et al.*, 2014; Josko and Oleszczuk 2014).

Cytotoxicity

During the risk evaluation of human health, various human cells have been used to explore adverse effects based on the various exposure routes of nano-TiO₂, such as bronchial epithelial cells (Bhattacharya et al., 2009), keratinocytes (Kocbek et al., 2010), lymphocytes (Ghosh et al., 2010), macrophage (Morishige et al., 2010) and hepatocyte (Sha et al., 2014). During the risk assessment of NMs in vitro, nano-TiO2 interacted with cultured cells directly. The changes in cell membrane integrity (Vevers and Jha 2008), mitochondrial activity (Di Virgilio et al., 2010), apoptosis marker (Veranth et al., 2007) and DNA structure (Ghosh et al., 2010) were detected after NMs exposure. Consequently, nano-TiO2 killed cells in dose- and time-dependent manners (Sha et al., 2011), through the generation of reactive oxygen species (ROS) (Bhattacharya et al., 2009), lysosomal membrane destabilization and lipid peroxidation (Hussain et al., 2010), changing the cellular morphology (Goncalves et al., 2010), cell-matrix adhesion (Fujita et al., 2009) and gene expression (Park et al., 2009), etc. Furthermore, genotoxicity, including the DNA damage, the formation of DNA adducts and the changes in the cell cycle, has been observed after NMs incubation (Falck et al., 2009; Kocbek et al., 2010).

In vitro test plays important roles in assessing the adverse effect of NMs (Corsi *et al.*, 2003), which can provide an initial estimation of the toxicity *in vivo* and facilitate a mechanistic understanding of the *in vivo* toxicity (Table 1). Nevertheless, testing may lead to false results of cytotoxicity owing to the different cell microenvironments *in vitro* as compared with its native counterpart (Heng *et al.*, 2010), thus an *in vivo* test based on various animals (e.g. mice and rat) has been used to better mimic the response of the human body to NMs (Tables 2, 3).

During animal experiments, NMs can enter mice or rat body through inhalation (Hougaard *et al.*, 2010), intraperitoneal (i.p.) administration (Chen *et al.*, 2009), i.g. administration (Cui *et al.*, 2011) and intratracheal (i.t.) instillation (Silva *et al.*, 2013). Nano-TiO₂ interacted with biological components of the body, leading to biodistribution in different tissues during/after exposure (Wu *et al.*, 2009), such as the lung (Hougaard *et al.*, 2010), brain (Hu *et al.*, 2010), liver (Cui *et al.*, 2011), heart (Sha *et al.*, 2013) and kidney (Chen *et al.*, 2009). Regardless of which administration is used to expose animals, the lung is a major injured organ. NMs can migrate to the interstitium of the lung burden, leading to the cell structure damage, alveolar macrophage and microvascular dysfunction, neutrophil infiltration, lung inflammation, thrombosis, alveolar



Figure 1. Schematic of titanium dioxide (TiO_2) nanotoxicity at the cellular level. Cell- TiO_2 nanomaterials (NMs) interactions can induce the membrane destabilization, Reactive oxygen species (ROS) production and lipid peroxidation during the penetration process of NMs (Davis *et al.*, 2010). The accumulation of ROS can perturb the antioxidant defense responses and induce OS condition in cells. Intracellular nano- TiO_2 can affect the normal function of macromolecules through protein adsorption, encumbering various signaling pathways, and inserting themselves into DNA base pairs (Li *et al.*, 2010; Drevet *et al.*, 2012; Kim *et al.*, 2013). After distributing to different organelles, the change of mitochondria membrane permeability, membrane potential, DNA conformation, and cell cycle were shown with the destruction of cytoskeleton integrity, DNA double-strand damage, and chromosomal segregation (Tarantola *et al.*, 2009; Trouiller *et al.*, 2009; Li *et al.*, 2010).

septal thickening, etc (Hougaard et al., 2010; Moon et al., 2010; Silva et al., 2013). For the brain, Nano-TiO₂ impaired spatial recognition memory behaviors of mice through disturbing the homeostasis of neurotransmitter system and trace elements (Hu et al., 2010), and induced the changes of the gene expression associated with brain development, neuronal apoptosis, oxidative stress (OS), and inflammation after long-term exposure (Shimizu et al., 2009). Nano-TiO₂ accumulated in the liver and led to focal hepatocytes apoptosis and hepatic dysfunction after i.g. administration (Hu et al., 2010), after histopathological changes, such as loss of the sinusoid space, hydropic degeneration with minor fatty change, and inflammation (Chen et al., 2009; Cui et al., 2011). Saccus lymphaticus, proteinic liquids and dilatation in the renal tubular. and swelling in the renal glomerulus were shown after the accumulation of nano-TiO₂ in the kidney (Chen et al., 2009; Wu et al., 2009). More notably, nano-TiO₂, which came from the i.g. administration, can distribute into the reproductive system, leading to ovary injury and fertility reduction through the regulation of inflammation/follicular atresia-related cytokine expressions (Zhao et al., 2013). Further, inhaled nano-TiO₂ enhanced prepulse inhibition of exposed female offspring and altered their neurobehavior (Hougaard et al., 2010).

Ecotoxicity

Nano-TiO₂ can induce ecotoxicity owing to industrial processes and consumer goods, such as the use of food-grade nano-TiO₂ (Yang *et al.*, 2014) and wastewater disinfectant (Yang *et al.*, 2013). After NMs spread to water, nano-TiO₂ showed adverse effects on aquatic animals (Table 4), including high mortality of medaka (Ma *et al.*, 2012), changes in gene expressions about the ribosome structure in zebrafish gills (Griffitt *et al.*, 2009), severe growth retardation, mortality and reproductive defects in daphnia magna (Zhu *et al.*, 2010) and developmental neurotoxicity in zebrafish through phototoxicity after UV radiation (Wang *et al.*, 2014). For hydrophyte, nano-TiO₂ affected the functions of photosynthesis-related genes and increased ROS levels through UV radiation (Miller *et al.*, 2012; Simon *et al.*, 2013).

In addition, nano-TiO₂ can interact with other metals (e.g. Pb) and modify their bioavailability and toxicity in aquatic environments. Nano-TiO₂ exhibited synergistic effects with heavy metal ions (e.g. Zn^{2+}) to increase the phytotoxicity on the photosynthetic capacity and algae survival rate (Tang *et al.*, 2013). Compared with Pb treatment alone, the bioconcentration and toxicity of Pb were significantly increased when combined with nano-TiO₂ (0.1 mg l⁻¹) in zebrafish larvae, which led to the disruption of thyroid endocrine and the neuronal system through affecting gene transcription (Sun *et al.*, 2014).

Nano-TiO₂ present in the soil also affects the bacteria (Table 5) and terrestrial plants (Table 6). For soil bacterial communities, microbial biomass and diversity were reduced after nano-TiO₂ exposure, which may alter the composition and function of the bacterial community in ecosystems (Ge *et al.*, 2011; Ge *et al.*, 2013; Binh *et al.*, 2014). Nano-TiO₂ can be absorbed from soil by terrestrial plants as they grow, accumulating in stems, leaves and fruits through plant metabolic systems, and affecting the germination, root elongation of seed and seedlings (Song *et al.*, 2013a, 2013b). Certainly, to avoid the effects of the soil's properties and various components in it, hydrophonic culture was used in nanotoxicity assessment of plants, which suggested the more toxic be in water than in soil (Josko and Oleszczuk, 2014).

Considering the top position of people in the food chain, the ecological environment is necessary for humans (D'Agata *et al.*, 2014). As a result, ecotoxicity, induced by nano-TiO₂ exposure in the water and soil, will translate to cytotoxicity after the interaction between people and the surroundings.

Potential mechanisms

In vitro, direct cell-TiO₂ NMs interactions can stimulate adverse effects, inducing ROS production, mitochondrial dysfunction, DNA damage, cell apoptosis, etc (Di Virgilio *et al.*, 2010; Ghosh *et al.*, 2010). *In vivo*, nano-TiO₂ can penetrate into different tissues (e.g. lung, liver, spleen and kidney), and affect the normal function of cells in tissues after complicated biodistribution, metabolism and clearance (Chen *et al.*, 2009; Wu *et al.*, 2009). The mechanisms of nanotoxicity at the cellular level are shown schematically in Fig. 1.

ROS generation and OS

As one of the widely accepted toxic mechanisms, the enhanced generation of ROS is shown after NMs exposure (Gonzalez *et al.*, 2008). The adverse effects of ROS production on cells have been reported for various NMs, including nano-TiO₂ (Long *et al.*, 2007). The accumulation of ROS can perturb the biological antioxidant defense responses (oxidant/antioxidant balance) (Xia *et al.*, 2004; Foster *et al.*, 2006). Owing to the increased ROS levels, NMs can deplete GSH (Sha *et al.*, 2013), dopamine (DA), dihydroxyphenylacetic acid (DOPAC) and homovanillic acid (HVA) in a dose-dependent manner (Hussain *et al.*, 2006). The ROS generated by NMs exposure can rapidly damage biological targets (e.g. bronchial epithelial cells) (Bhattacharya *et al.*, 2009) and trigger the apoptotic process through activating cytosolic caspase-3 and chromatin condensation (Park *et al.*, 2008).

The imbalance between the production and detoxification of reactive oxygen will induce a state of OS in cells (Finkel and Holbrook, 2000). In the OS model, antioxidant defense, proinflammatory effects and cytotoxicity are three classes of OS aggravated by the incremental cellular responses under the production of ROS (Nel et al., 2006; Meng et al., 2009). Based on the published literature, the transcriptional activation of the antioxidant response element in phase II enzyme promoters can mediate the homeostatic redox balance pathway (Xia et al., 2006). Then, mitogen-activated protein kinase (MAPK), AP-1 and nuclear factor kappa-B (NF-kB) signaling cascades can activate cytokine and chemokine expressions (Xiao et al., 2003). The increased levels of transcription factor NF-kB and oxidant-dependent inflammatory signaling can induce the cellular inflammatory response when lungs are exposed to nano-TiO₂ (Moon et al., 2010). Nano-TiO₂ can induce an increased level of IL-8 mRNA through both transcriptional and post-transcriptional pathways (Park et al., 2008). Finally, mitochondrial perturbation mediates the toxic OS and activates cytotoxic cell death. Exposure of the cultured cells to NMs can induce the activity of OS-related genes, such as catalase and thioredoxin reductase (Park et al., 2008), and cell apoptosis was triggered by the activation of caspase-8/Bid and the mitochondrial signaling pathways (Zhao et al., 2009).

Organelle dysfunction

NMs are similar in size to many biological molecules and structures (~0.1 to 100 nm), enabling NMs to easily enter cells, organelles and functional biomolecular structures and interact with vital biological systems (Fischer and Chan, 2007). Most importantly, intracellular NMs may have observable toxicity because they are not bound to the cell membrane and interact with the macromolecule directly (Geiser *et al.*, 2005). For example, the adsorption of proteins onto the NMs surface can also change protein conformation (Lynch *et al.*, 2006). Consequently, the abnormal behavior of

protein can encumber various signaling pathways, leading further to cell death (Berntsen *et al.*, 2010).

Mitochondria supply energy to cells and keep their survival (Calabrese *et al.*, 2001). After nano-TiO₂ exposure, mitochondrial membrane permeability experience significant change, which may lead to the release of cytochrome c from mitochondria to the cytosol (Zhao *et al.*, 2009). Mitochondrial perturbation, such as the loss of mitochondrial membrane potential and mitochondrial damage, leads to cell death (Hussain *et al.*, 2005).

Cytoskeleton (CSK) is a compact network of structural proteins and crucially important to intracellular transport and cellular division (Bursac *et al.*, 2005). An *in vitro* study suggested that NMs can affect cell viability through the destruction of CSK integrity (Tarantola *et al.*, 2009). During the process of cell penetration, NMs can change the normal function of the cytoskeletal network via caveolin, endocytosis and adhesive interactions (Rothen-Rutishauser 2009). After neonatal rat ventricular myocytes have been exposed to nano-TiO₂, the myofibrils showed a less organized structure than the non-exposed cells. TEM examinations revealed that nano-TiO₂- exposed cells exhibit thick electron dense Z-lines, compared with the control (Helfenstein *et al.*, 2008).

For cell nucleus, DNA damage caused by exposure to NMs is one of the key reasons behind the development of cell death and tumor. After the exposure of the liver cell to anatase nano-TiO₂, which inserted themselves into DNA base pairs or binds to DNA nucleotide, resulting in the change of DNA conformation (Li *et al.*, 2010). Nano-TiO₂ can cause the formation of gamma-H2AX foci and 8-hydroxy-2'-deoxyguanosine, inducing oxidative DNA double-strand damage in a mice model (Trouiller *et al.*, 2009). NMs also disturb cell-cycle progression and lead to further chromosomal segregation and cell transformation. After long-term exposure to nano-TiO₂, cultured fibroblast cells demonstrated increasing numbers of multinucleated cells and micronucleus, G2/M delay and slower cell division (Huang *et al.*, 2009).

Conclusions

In this paper, we detail the current knowledge of the adverse effects induced by nano-TiO₂. Based on the abundance of risk studies on nano-TiO₂, the detailed information, such as, evaluative object, physico-chemical properties of nano-TiO₂, exposure concentration, exposure duration and toxicity, were listed in Tables 1–6. However, there are limited informational values for nanotoxicity assessment in many published reports, and the lack of toxicity standard limits the deeper understanding of potential toxicity.

First, the physico-chemical properties of nano-TiO₂ were not described adequately. The crystal structure, shape, size, treatment media, zeta potential, aggregation and agglomeration tendency, and modification of nano-TiO₂ as the essential information were collected in our tables. Various forms of nano-TiO₂ led to different toxicity, just like the different responses of bronchial epithelial cells (BEAS-2B) as a model (Bhattacharya *et al.*, 2009; Falck *et al.*, 2009). Unfortunately, it is hard to compare the different adverse effects between various nano-TiO₂ mainly because of missing the essential description of nano-TiO₂ used.

Second, extremely high doses of nano-TiO₂ have been used in acute toxicity research, which may be impractical (Chen *et al.*, 2009). Animal experiments have enabled researchers to perform, replicate and quantify the adverse factors detected from *in vitro* studies, such as ROS, signal pathway activity, OS, cell cycle and inflammation (Poland *et al.*, 2008; Ryman-Rasmussen *et al.*, 2009). However, the adverse effects, as shown *in vitro* and *in vivo* after a

high dose of nano-TiO₂ exposure, are not adequate to provide an accurate prediction of human body responses.

Third, to take into account the difference between exposure routes using in vivo studies is also important to identify and quantify the nanotoxicity. For example, the increased usage of nano-TiO₂ in cosmetics suggests that the skin is a route for human to get exposed to NMs. Wu et al. (2009) found that nano-TiO₂ penetrated through the skin into different tissues and induced diverse pathological lesions after 60 days exposure in hairless mice. In contrast, Adachi et al. (2013) found no evidence of nano-TiO₂ skin penetration after 56 days exposure of NMs emulsion to the hairless rat. To avoid the difference of NMs and species, the meaningful risk assessments should be the epidemiological study of heavily-exposed populations, which can reflect the real effects through the respiratory system after nano-TiO₂ exposure (Warheit, 2013). For example, based on the available data and hygiene standards, Swidwinska-Gajewska et al. (2014) concluded that 0.3 mg m⁻³ nano-TiO₂ is the maximum admissible concentration in the workplace in Poland. In addition, more experiments on inhalation exposure are needed to measure the relevance of risk effects for human.

Based on the nanotoxicity assessment studies, nano-TiO₂ induced health risks, such as cytotoxicity, ecotoxicity, phototoxicity and phytotoxicity, as caused by the generation ROS and OS, organelle dysfunction, etc. Actually, the life cycles of nano-TiO₂ can be determined with the development of the nano-risk framework, the adverse effect of nano-TiO₂ will effectively decrease with the improvement of strict preventive procedures if consistent definition and standard of low toxicity is established. Better understanding of the toxicity of nano-TiO₂ holds promise for strict prevention of exposure to NMs, and will facilitate the use of the superior power of these double-edged weapons.

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Conflict of interest

None of the authors has any conflicts of interest related to this manuscript.

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